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The Response of Macrophytes to Nutrients and Implications for the Control of Phytoplankton Blooms in East Taihu Lake, China

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Abstract

When macrophytes are growing in the eutrophicated aquatic ecosystem, the vegetation induces important effects to the water quality and phytoplankton concentrations in the water which affected by macroscopic physical, chemical and biological processes and the effects are the results of direct and indirect interactions of the aquatic plants and water body. The interactions between macrophytes, nutrients and phytoplankton blooms were examined in the water and sediments of a shallow, eutrophic and typical East Taihu, China. The importance of macrophytes as a sink for nutrients, and the inhibitory effect of macrophytes on phytoplankton bloom potential were assessed through three different seasons. Luxuriant aquatic plants growth in this system led to decrease available nutrients for phytoplankton and prevented bloom development. Uptake of N and P by aquatic plants accounted for a major portion of the observed N and P loss from the water column and sediments. Luxury uptake of N and P were indicated by high biomass and tissues N and P concentrations, indicating the capacity of macrophytes to act as a nutrient sink in midsummer. Dissolved inorganic N (DIN) and soluble reactive phosphorous (SRP) in water were reduced in midsummer in the presence of macrophytes. The use of macrophytes to reduce the nutrients in water system and thereby inhibit freshwater phytoplankton blooms should be considered as an effective management strategy in shallow eutrophicated lakes. As aquatic macrophytes also develops considerable indirect effects that could have a vital impact than the direct uptake the nutrients into the plant biomass.

Keywords: Eutrophication; Macrophytes; Nutrients; Algae blooms; East Taihu Lake

Introduction

Submerged macrophytes are an important biotic component in freshwater ecosystems worldwide [1]. They provide refuge and food for various animals [2] and exert a strong influence on the water physical and chemical properties [3]. Macrophytes alter (generally increase) biodiversity in aquatic habitats [4]. They also provide shelter for zooplankton and young fish, reduce nutrient levels, serve as a habitat for macro-invertebrates [1,4]. Some field studies have demonstrated that the "clearing effect" increased with the macrophytes density and spatial extension of stands, but it was only restricted to a short distance outside the vegetation [5,6].

Freshwater phytoplankton blooms (mainly cyanobacterial) have become an increasingly problematic water quality issue worldwide [7-9]. They represent a health threat to domestic animals and human consumers of affected waters [10,11]. Blooms are primarily caused by excessive loading of nutrients [8] and global warming appears to enhance bloom potentials [12]. In recent years, periodic and widespread phytoplankton blooms have proliferated in the Taihu, China and this phenomenon was particularly serious in Taihu Lake in 2007 [13,14].

Submerged macrophytes can play an important role in the control of phytoplankton within enclosed shallow water bodies [15,16], and a possible negative feedback exists between shading provided by macrophytes and phytoplankton bloom development [17]. Enhanced nitrogen (N) and phosphorus (P) uptake and accumulation by summer biomass buildup of macrophytes leads to N limitation of the phytoplankton [15,18]. We investigated nutrient and macrophyte interactions with respect to their impacts on phytoplankton bloom potentials and water quality in East Lake Taihu, China (referred to as East Taihu). The lake has distinct regions where macrophytes form an important fraction of primary producers. In the relatively clear waters of East Taihu, macrophytes tend to dominate. Conversely, primary production in relatively turbid North Taihu is dominated by phytoplankton. So the purpose of the study was to investigate the role macrophytes played in nutrient cycling in the system by examining the partitioning of N and P between the macrophytes, water and sediment in East Taihu from March to December in 2009.

Material and Methods

Study area

The Eastern portion of Taihu is largely comprised of a large (130 km²), shallow (mean depth 1.0 m) Bay, which was called Eastern Taihu. Submerged macrophytes have historically flourished in the lake region [19]. This region is also an important water source for the City of Shanghai [19]. The entire Taihu is currently eutrophic, with total phosphorus (TP) and total nitrogen (TN) concentrations of the water column rarely falling below the OECD threshold value of 0.02 mg/L TP and 0.2 mg/L TN for eutrophic lakes [20].

Three sampling sites (1#, 2#, 3#) in eastern Taihu were selected for transects according to prior results from the Taihu Monitoring

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Program Stations (Figure 1). The three sites of 1# (E:120.41428°, N:30.98694°), 2# (E:120.4342°, N:31.03294°) and 3# (E:120.49765°, N:31.08131°) were located at the east suburb channel, near commercial crab culture operations, and the main channel of open water region, respectively. Sites were sampled in the March, August and December of 2009. These months represented the germination, maturation and senescence periods for aquatic macrophytes.

Sampling and analytical methods

Six quadrats were located along each transect crossing the studying area. Every sampling site had two quadrats. Water quality data and water samples were collected from all quadrats before collection of submerged macrophytes and sediment samples were undertaken to prevent contamination of the water with sediment. Secchi depth, temperature and dissolved oxygen (DO) were measured on sites. The suspended solids (SS), pH and EC of the water were measured in the laboratory.

Water samples were obtained using a water sampler deployed three times at different depths. Sub-samples were collected for chemical analyses. Surface sediment per quadrat was sampled and then frozen immediately in dry ice. The samples were analyzed for sediment chemical composition on the pooled samples of the top 2 cm of sediments. This is considered to be the portion of sediment where nutrients, P in particular, are enriched [21]. Lastly, the submerged macrophytes were collected from each quadrat, with either a submerged 0.28 m² quadrat by hand or a circular rake for very small plants [22] or large plants, respectively. The plants were then dried at 60°C for 24 h and weighed.

Water samples were analyzed for total phosphorus (TP), total dissolved phosphorus (TDP), soluble reactive phosphorous (SRP), total nitrogen (TN), total dissolved nitrogen (TDN), ammonium (NH₄⁺-N), and nitrogen oxides (NO_x-N). Concentrations of TDN, NH₄-N and NO_x-N were determined colorimetrically on a Skalar Autoanalyser (Skalar-SA 3000/5000, Netherlands). TN, TDN were pretreated by digestion with potassium sulphite. TDP and TP were determined after perchloric acid digest, and SRP was determined using the ascorbic acid:molybdate method [23]. Chlorophyll *a* was measured with hotethanol extraction method as the phytoplankton concentration in the sampling water [24,25]. Sediment TP was measured colorimetrically of Murphy and Riley [26].

Results

Environmental conditions in the East Taihu

Surface water temperature increased from a March mean of 13°C to an August maximum average of 32.0°C (Table 1), and then declined

to a December mean of 9.0°C. Bottom temperatures were similar to the surface water in spring, increasing more gradually over the summer time. The trend in surface DO shows a decline from March to August, then a rise to December, with some variability among sampling sites. The water pH and EC varied relatively little, except for some decrease in the aquatic macrophytes maturation period. Water Secchi depth of the 1# site was 0.15 m in March, and at the same time it were 0.90 m and 0.95 m at 2# site and 3# site, respectively. But the Secchi depth of 1# site reached 1.60 m in August, it was higher than that the other two sites and the water Secchi depth changed along with the season and macrophytes growth.

Nutrients changes in water and sediment

The selected three sampling sites exhibited similar water depth profiles and fluctuations (Figure 2). However, the three sites showed some variability among different forms of nutrients in the water. They had the higher concentration in March at 1# site, and then dropped to the minimum in August, but then increased again in December. The other sites showed an inverse pattern. Dissolved nitrogen and phosphorus concentrations in water column were sufficiently high to be available to macrophytes and phytoplankton. Sites 2# and 3# had





near the crab culture and the main channel of open water region, respectively).

Sampling site 1#					2#			3#		
Sampling time	Mar	Aug	Dec	Mar	Aug	Dec	Mar	Aug	Dec	
Secchi depth (m)	0.15	1.60	1.00	0.95	1.22	0.90	0.45	0.85	1.30	
Water depth (m)	1.00	1.60	1.50	0.95	1.22	0.90	1.00	1.60	1.40	
Temperature (m)	13.0	32.0	9.0	13.0	32.0	9.0	13.0	32.0	9.0	
рН	7.92	8.33	8.02	8.72	7.80	7.79	8.09	7.85	7.87	
EC (µs/cm)	560.0	520.0	460.0	420	500.0	520.0	510.0	420.0	500.0	
DO (mg/l)	10.10	9.85	4.50	7.56	4.87	9.70	8.30	7.65	9.60	
SS (mg/l)	118.64	2.16	7.77	0.80	2.61	9.30	20.88	6.56	8.93	

Table 1: The water environmental characteristic in East Lake Taihu in 2008

(Integrated samples of the water column were obtained using a water sampler deployed three times at different depths. Then water samples were brought back to the laboratory with the ice-bag protected and analyzed as soon as possible. The each sampling time was chosen at the same 9:00 am; and all values presented are mean values).

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the lower concentration of the TDN and TDP, and then increased from March to August. During the period from August to December, only the concentration of TDP in 1# sampling site rose (Figure 3). The PO_4^{3-} -P and NO_3^{-} -N in water appeared to be directly available to macrophytes and phytoplankton. Though the NO_3^{-} -N was higher in March, it decreased quickly in summer.

The TN and TP concentration in the sediments were similar to those in the water column. However, the 2# site had the higher concentration in March, and it was the time of the germination period of aquatic macrophytes, after which the nutrient concentrations declined during the summer (Figure 4). Nutrient accumulations appeared largely attributable to macrophytes decomposition.

Macrophytes and phytoplankton in the water column

The three sampling sites showed similar environmental conditions in the water and the macrophytes occurred in shallow regions to a maximum depth of 1.60 m in August and a maximum depth of 0.95 m in March. Macrophytes community composition also varied with sampling times and sites. The main species of the macrophytes included: *Potamogeton malaianus Miq.; Vallisneria natans L.; Elodea nuttallii; Hydrilla verticillata Royle; Ceratophyllum demersum* with other species comprising only a small percentage of the total composition (data not shown).

Site 1# is located at the east suburb channel, the flood discharge channel of Taihu and main water supply for the City of Shanghai. The macrophytes could obtain amounts of nutrients and grow well as the water velocity became lower at the 1# site (Figure 5). The 2# site had the highest biomass of aquatic plants during the selected sampling sites in March. But the 1# site had the highest biomass among three sampling sites in summer (fresh weight reaching 2,255.48 g/m²). The 3# site had low biomass relatively except in August. Though the species composition and the macrophyte biomass showed little difference during the sampling time, in general there was an increase in aquatic plant biomass at sampling sites except some decreases at the 3# sampling site.

Phytoplankton biomass (dominated by cyanobacterial), estimated as Chl-a concentrations, varied among sampling times and sites (Figure 5). The concentration of phytoplankton biomass at 1# site was higher than that at the others sampling sites in March, but it was the

Discussions

Conditions within the water column

The increase in water temperature in early summer at the germination period of aquatic macrophytes would favor growth of







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these plants. The early summer increase in biomass of photosynthetic plants would explain the increase DO concentrations in water in March. Differential photosynthetic use of carbon dioxide would also explain the fluctuation in pH at the three sampling sites. Higher water temperatures could also cause a decrease in the oxygen solubility in the water and therefore the August DO concentrations were reduced at all sampling sites (Table 1). As some macrophytes decomposed during this senescent period, DO consumption would also increase. As mentioned above, the macrophytes grew most vigorously from March to October, as temperature favored the aquatic plants and phytoplankton growth in this period. The results of Wang [27] indicate that water temperature and total phosphorus (TP) played dominant roles in controlling phytoplankton growth dynamics in most seasons; COD (chemical oxygen demand) and BOD (biological oxygen demand) presented significant positive relationships with phytoplankton biomass in spring, summer and autumn.

Macrophytes and nutrients

N and P values in the Taihu sediment are significantly different from those found in East Taihu sediments [28-30]. TN and TP concentration in the water was higher in 2# site in March and then decreased through the summer period. As in the water column, the TN in sediment also had higher content in December than that in August, although TP varied little during this period. The N and P concentration changes in the sediments indicate that nutrient losses from the top 2 cm of sediment were similar to the nutrient gains of the macrophyte community (Figure 5), of which the major species was the submerged macrophyte *Potamogeton malaianus*. The N data showed an increase in macrophyte N which was not matched by loss from the sediment, a phenomenon that has been observed by others [31,32]. TDN decreases in water column, although considerable (Figure 3). The additional N may have been derived from deeper sediments or from western Taihu water, although P data do not support this (Figure 3).

The macrophytes biomass progressively increased from March to August (Figure 5), a period when temperature and irradiance are favorable for macrophyte growth. Macrophyte species diversity was also enhanced during the same period (data not shown). Thriving aquatic plants during this period will optimally adsorb dissolved nutrients from the water, and increase the depth of the photic zone (Table 1). In a word, these effects would promote a positive-feedback, environment friendly ecosystem [33].

The Inhibition Effects of Macrophytes on Phytoplankton

The use of macrophytes for reducing N and P availability and thereby reducing phytoplankton blooms potential has been used in diverse aquatic ecosystems [18,34,35]. Possible inhibition of phytoplankton by allelochemicals released by submerged macrophytes has been proposed as one of the mechanisms that contribute to the maintenance of clear-water states in shallow lakes [36,37]. Although direct proof of allelopathy remains elusive, several authors have suggested possible involvement of allelopathy to explain phytoplankton successional patterns in whole-lake studies of vegetated, shallow lakes [38,39]. During this survey, the Chl-a concentration was low at the sampling sites except for 1# in March, when it was somewhat higher (Figure 5). The reason have been that this sample was taken from a site that was influenced by open Taihu water which passed through the 1# site, which was the main flood discharge channel of the Taihu. Following the summer macrophyte growth period, phytoplankton concentrations rapidly decreased. The macrophytes decayed at higher temperature in summer and caused oxygen deficiency in water column

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at the 2# site, as proven the nutrients concentration variety (e.g. higher concentration of $PO_4^{3-}P$, TDP in water). The apparent inhibitory effects of macrophytes on phytoplankton are consistent with the other results [40-42]. Another explanation is salinity (conductivity) control of nutrient uptake by macrophytes [43], as the conductivity changed during surveying time (ranged from 560 to 42 µs/cm during the March to August time). Decreased salinity would promote uptake of N and P by these macrophytes, resulting in decreased availability of nutrients for phytoplankton.

Conclusions

Aquatic macrophytes have been shown to be a significant sink for nutrients and they increased clarity in East Taihu. Higher summer macrophyte biomass was responsible for uptaking large amounts of N and P. Macrophytes incorporation of N and P accounted for most of the observed nutrient loss from the sediment and water column. Surficial sediment N loss did not account for all the macrophyte N gain. Therefore other N sources, including cyanobacterial N₂ fixation and atmospheric N deposition, are postulated. Water column TDN and NO₃-N significantly decreased during the *P. crispus* growth period. Decreased salinity (conductivity) maybe have also promoted N and P incorporation by macrophytes and hence the reduction of available nutrients for phytoplankton, explaining a reduction in phytoplankton biomass and blooms during the summer. The inhibitory effect of macrophytes on phytoplankton blooms has significant ramifications for management of phytoplankton blooms.

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Author Disclosure Statement

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References

- Jeppesen E, Søndergaard M, Christoffersen K (1998) The structuring role of submerged macrophytes in lakes. Springer, New York, USA.
- 2. Proctor VW (1999) Charophytivorie, playas y papalotes, a local paradigm of

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global relevance. Australian Journal of Botany 47: 399-406.

- Sand-Jensen K, Frost-Christensen H (1998) Photosynthesis of amphibious and obligately submerged plants in CO₂-rich lowland streams. Oecologia 117: 31-39.
- Scheffer M, Jeppesen E (1998) Alternative stable states. In: Jeppesen E, Søndergaard M, Christoffersen K (eds.) The Structuring Role of Submerged Macrophytes in Lakes. Ecological Studies 131: 397-406.
- James WF, Barko JW, Butler MG (2004) Shear stress and sediment resuspension in relation to submersed macrophyte biomass. Hydrobiologia 515: 181-191.
- Horppila J, Nurminen L (2005) Effects of different macrophyte growth forms on sediment and P resuspension in a shallow lake. Hydrobiologia 545: 167-175.
- 7. Reynolds CS (1987) Cyanobacterial water blooms. Adv Bot Res 13: 67-143.
- Paerl HW (1988) Nuisance phytoplankton blooms in coastal, estuarine, and inland waters. Limnol Oceanogr. 33: 823-847.
- Paerl HW, III Fulton RS (2006) Ecology of harmful cyanobacteria. In: Graneli E, Turner J (eds.). Ecology of Harmful Marine Algae. Springer-Verlag, Berlin.
- Falconer IR, Burch MD, Steffensen DA, Choice M, Coverdale OR (1994) Toxicity of the blue-green alga Microcystis *aeruginosa* in drinking water to growing pigs, as an animal model for human injury and risk assessment. Environmental Toxicology and Water Quality 9: 131-139.
- 11. Chorus I, Bartram J (1999) Toxic cyanobacteria in water: A guide to their public health consequences, monitoring and management, London.
- 12. Paerl HW, Huisman J (2008) Blooms like it hot. Science 320: 57-58.
- Guo L (2007) Doing Battle with the Green Monster of Taihu Lake. Science 317: 1166.
- 14. Song LR, Chen W, Peng L, Wan N, Gan N, et al. (2007) Distribution and bioaccumulation of microcystins in water columns: A systematic investigation into the environmental fate and the risks associated with microcystins in Meiliang Bay, Lake Taihu. Water Research 41: 2853-2864.
- Ozimek T, Gulati RD, Donk EV (1990) Can macrophytes be useful in biomanipulation of lakes? The Lake Zwemlust example. Hydrobiologia 61: 399-407.
- 16. Jackson HO, Starrett WC (1959) Turbidity and sedimentation as Lake Chautauqua, Illinois. J.Wild.Manage. 23: 157-168.
- Scheffer M, Rinaldi S, Gragnani A, Mur LR, van Nes EH (1997) On the dominance of filamentous cyanobacteria in shallow, turbid lakes. Ecology 78: 272-282.
- Donk EV, Gulati RD (1995) Transition of a lake to turbid state six years after biomanipulation: Mechanisms and pathways. Water Science and Technology 32: 197-206.
- Huang YP (2001) The Water Environmental and Contamination Control of Taihu Lake. Beijing, Science Press, China.
- 20. Ryding SO, Rast W (1989) The Control of Eutrophication of Lakes and Reservoirs. UNESCO. Paris and Parthenon.
- Howard-Williams C (1981) Studies on the ability of a Potamogeton pectinatus community to remove dissolved nitrogen and phosphorus compounds from lake water. J. Appl. Ecol. 18: 619-637.
- Marshall TR, Lee PF (1994) An inexpensive and lightweight sampler for the rapid collection of aquatic macrophytes. J. Aquat. Plant Manage. 32: 77-79.
- Major GA, Dal Pont G, Klye J, Newell B (1972) Laboratory techniques in marine chemistry - a manual. CSIRO Division of Fisheries and Oceanography.
- Jespersen AM, Christofersen K (1987) Measurements of chlorophyll-a from phytoplankton using ethanol as extraction solvent. Arch Hydrobiologia 109: 445-454.
- Chen YW, Chen KN, Hu YH (2006) Discussion on possible error for phytoplankton Chl-a concentration analysis using hot-ethanol extraction method [J]. J.Lake Sci. 18: 550-552.
- Murphy J, Riley JP (1962) A modified single solution method for the determination of phosphate in natural waters. Anal. Chim. Acta 27: 31-36.
- 27. Wang XL, Lu YL, He GZ, Han JY, Wang TY (2007) Exploration of relationships between phytoplankton biomass and related environmental variables using

multivariate statistic analysis in a eutrophic shallow lake: A 5-year study. Journal of Environmental Sciences 19: 920-927.

- 28. Fan CX, Zhang L, Qin BQ, Hu W, Gao G, et al. (2004) Migration mechanism of biogenic elements and their quantification on the sediment-water interface of Lake Taihu:I.Spatial variation of the ammonium release rates and its source and sink fluxes. J. Lake Sciences 16: 10-20.
- 29. Qin BQ, Hu WP, Gao G, Luo L, Zhang J (2004) Dynamics of sediment resuspension and the conceptual schema of nutrient release in the large shallow Lake Taihu, China. Chinese Science Bulletin 49: 54-64.
- Xu DL, Lei ZX, Wang HJ, Han BP, Liu ZW (2007) Distribution of Phosphorus in Sediments of Onshore Reed Areas of Lake Taihu. J.China Univ Mining & Technol 17: 557-561.
- Chen RL, Barko JW (1988) Effects of freshwater macrophytes on sediment chemistry. Journal of Freshwater Ecology 4: 279-289.
- Wilma JV (2001) Nutrient partitioning in the upper Canning River, Western Australia, and implications for the control of cyanobacterial blooms using salinity. Ecological Engineering 16: 359-371.
- 33. Scheffer M (2001) Ecology of Shallow Lakes. Kluwer Academic Publishers.
- 34. Stuart FM, Robert TW (1996) Grazing by black swans (*Cygnus atratus Latham*), physical factors, and the growth and loss of aquatic vegetation in a shallow lake. Aquatic Botany 55: 205-215.
- Cassandra SJ, John WE, Keith H (2006) Responses of three invasive aquatic macrophytes to nutrient enrichment do not explain their observed field displacements. Aquatic Botany 84: 347-353.
- Daniela E, Elisabeth MG (2006) Allelopathic activity of *Elodea canadensis* and *Elodea nuttallii* against epiphytes and phytoplankton. Aquatic Botany 85: 203-211.
- Mulderij G, Mooij WM, Smolders AJP, Van Donk E (2005) Allelopathic inhibition of phytoplankton by exudates from *Stratiotes aloides*. Aquatic Botany 82: 284-296.
- Blindow I, Hargeby A, Andersson G (2002) Seasonal changes of mechanisms maintaining clear water in a shallow lake with abundant Chara vegetation. Aquatic Botany 72: 315.
- 39. Lombardo P (2005) Applicability of littoral food-web biomanipulation for lake management purposes: Snails, macrophytes, and water transparency in northeast Ohio shallow lakes. Lake and Reservoir Management 21: 186-202.
- Canfield DE, Shireman JV, Colle DE, Haller WT, Watkins II CE (1984) Prediction of chlorophyll a concentrations in Florida lakes: Importance of aquatic macrophytes. Canadian Journal of Fisheries and Aquatic Sciences 41: 497.
- 41. Meijer ML, Jeppesen E, Van Donk E, Moss B, Scheffer M, et al. (1994) Longterm responses to fish-stock reduction in small shallow lakes: Interpretation of five-year results of four biomanipulation cases in The Netherlands and Denmark. Hydrobiologia 275-276: 457-466.
- Mjelde M, Faafeng B (1997) Ceratophyllum *demersum* hampers phytoplankton development in some small Norwegian lakes over a wide range of phosphorus concentrations and geographical latitude. Freshwater Biology 37: 355-365.
- Moss B (1994) Brackish and freshwater shallow lakes-different systems or variations on the same theme? Hydrobiologia 275-276: 1-14.